# Influence of inorganic nutrients and dissolved organic matter on the growth of cyanobacteria *Microcystis aeruginosa* isolated from the Gulf of Riga

# Santa Purvina<sup>1</sup>\*, Christian Béchemin<sup>2</sup>, Maija Balode<sup>1</sup>, Daniel Grzebyk<sup>2</sup>, Serge Y. Maestrini<sup>2</sup>

<sup>1</sup>Latvian Institute of Aquatic Ecology, Daugavgrīvas 8, LV-1048, Riga, Latvia <sup>2</sup>Centre de Recherche en Ecologie Marine et Aquaculture de L'Houmeau (CNRS-IFREMER), B.P. 5, F-17137 L'Houmeau, France

\*Corresponding author, E-mail: santa.purvina@lhei.lv

# Abstract

The influence of nutrient additions in N-limiting, P-limiting and in DIN:DIP-balanced conditions with and without dissolved organic matter (DOM) of land origin was studied on the growth and toxin content of *Microcystis aeruginosa* isolated from the Gulf of Riga. The dissolved inorganic nitrogen (DIN) and dissolved inorganic phosphorus (DIP) supply in balanced ratio of DIN : DIP = 16 is the main determinative factor for *M. aeruginosa* development, promoting increases in cell biomass, chlorophyll *a* content and particulate organic matter. In medium enriched only by DOM, *M. aeruginosa* used 31 % of the available dissolved organic nitrogen and 53 % of the dissolved organic phosphorus. The experiment conditions did not invoke increment of toxicity. The experiments showed that the development of *M. aeruginosa* is promoted by the raised load of inorganic nutrients and dissolved organic matter to the brackish ecosystems of the Baltic Sea.

Key words: cyanobacteria, dissolved organic matter, *Microcystis aeruginosa*, nutrient limitation, the Gulf of Riga.

# Introduction

Nutrients, light and temperature are the most important factors for the development of marine phytoplankton. Human activities contribute to increased nutrient discharge to aquatic ecosystems from agriculture, municipal and industrial waste waters world wide. These anthropogenic nutrient sources have led to increased nutrient fluxes also in the Baltic Sea invoking eutrophication processes, followed by reduced water transparency, increased phytoplankton primary production, changes of planktonic species composition etc. (HELCOM 2007). Significant impact on nutrient input is favored also in terms of climate change. For the Baltic Sea ecosystems important changes are in the timing of the seasonal events: water discharge increment in winter in association with the reduction of spring runoff and shortening of the ice and snow cover period. Estimates indicate that river runoff may increase by up to 15 % averaged over the Baltic Sea catchment region, enlarging the export of dissolved organic matter to the brackish ecosystems (HELCOM 2007). Significant dependence of the river discharge on climate change has been also

found in Latvia (Kļaviņš, Rodionovs 2007).

Two bottom-up mechanisms have been invoked as major contributors in the increase of toxic algal blooms in last decades. From one side, human activities significantly increase the input of algal nitrogenous and phosphorus nutrients to estuarine and coastal waters, while the silicon concentration has remained constant or has even decreased in river loads to brackish estuaries (Rahm et al. 1996). It is assumed that in such conditions the proliferation of organisms having little or no requirement for silicon, such as flagellates and cyanobacteria, is favored (Humborg et al. 2000). On the other side coastal waters close to estuaries receive large quantities of dissolved organic matter (DOM), which may favor the growth of auxotrophic/photo-heterotrophic species versus that of autotrophic species (Paerl 1988).

Cyanobacteria have been responsible for the most harmful phytoplankton events in the Baltic Sea. Over the entire Gulf of Riga the potentially toxic cyanobacteria *Aphanizomenon flos-aquae*, *Nodularia spumigena*, *Snowella lacustris* and *Anabaena lemmermannii* can develop, whereas in the coastal areas common species are *Anabaena spiroides*, *Anabaena flos-aquae* and *Microcystis aeruginosa* (Balode, Purina 1996). The potentially toxic cyanobacteria *Microcystis aeruginosa* is mostly known to develop in freshwater, but it is also frequently observed in marine coastal waters (Kononen, Sellner 1995). Additionally, high concentration of *M. aeruginosa* biomass is transported to estuarine areas via rivers. In the coastal areas of the Gulf of Riga maximum development of *M. aeruginosa* continues from late June till September., the development usually occurs sporadically, but often are initiated blooms in the coastal zones (Balode, Purina 1996; Seppälä, Balode 1999). *M. aeruginosa* abundance is related with high nutrient concentrations (Rantajärvi et al. 1998). Studies on this species have intensified in the last decades mainly due to the discovery of potent hepatotoxic heptapeptides, called microcystins, shown to be deleterious to wild and domestic animals and also humans.

In natural water-bodies *M. aeruginosa* is found as several toxic and non-toxic genotypes. Whether a genotype can produce microcystins depends on the presence of intact genes from the microcystin synthetase gene complex (Kurmayer et.al. 2002). A seasonal succession of different *Microcystis* genotypes might often be a key mechanism determining microcystin concentrations in *Microcystis*-dominated lakes (Kardinaal et.al. 2007a) and other water-bodies. In nature several mechanisms might be involved in seasonal succession of toxic and non-toxic *Microcystis* strains: selective feeding by bivalves (Dionisio-Pires, Van Donk 2002) and zooplankton (Lürling 2003), *Microcystis* strains may differ in their buoyancy (Dunton, Walsby 2005) and competitive ability for light (Kardinaal et al. 2007b). The changes in nutrient concentrations and ratios in the environment affect not only phytoplankton community structure, but also algal physiology. Thus N-rich PSP toxins are synthesized during excess N and P-limitation but not during N-limitation (Granéli et al. 1998).

The estimated microcystin concentrations of *M. aeruginosa* isolated from the Gulf of Riga are low and range between 0.8 to 4.5 ng mg<sup>-1</sup> dry mass. The potentially toxic cyanobacteria *Nodularia spumigena* also can produce hepatotoxic toxin nodularin in high concentrations. However the observed toxicity of *Nodularia spumigena* in the Gulf of Riga is lower (nodularin 0.030 - 0.123 mg L<sup>-1</sup>) than in the Open Baltic Sea (0.03 - 1.35 ng mg<sup>-1</sup>); the toxin content of the Gulf of Riga *Nodularia spumigena* filaments is about 12 times higher than those from the open Baltic (Balode, unpublished data).

The aim of the present study was to test the effect of dissolved organic matter and raised inorganic nitrogen and phosphorus concentrations in different DIN : DIP ratios on the growth and toxicity of potentially toxic cyanobacteria *M. aeruginosa*, isolated from the coastal area of the Gulf of Riga, the Baltic Sea.

#### **Materials and methods**

The clone culture of *Microcystis aeruginosa* Kützing (MAGR-2) was isolated from the Gulf of Riga, Baltic Sea. The stock culture was grown in F medium of Guillard and Ryther (1962), the N : P ratio was adjusted to the Redfield value – 16 (1934). Iron was added in the form of Fe-EDTA.

The *M. aeruginosa* mother culture was obtained after centrifugation in sterile conic flasks. Supernatant containing the growth media was removed and cells were re-suspended in nutrient-free sterile seawater and left for 24 h in the culture room. Then, centrifugation and re-suspension were repeated, resulting in partly nutrient-depleted *M. aeruginosa* cells in a mother suspension free from dissolved nutrients. The culture was inoculated to experimental 5 L bottles, providing an initial concentration of *M. aeruginosa* 18.7 × 10<sup>6</sup> cells per liter. The cultures were grown at  $18 \pm 1$  °C, with a 16 h / 8 h light/dark period at 53 µmol m<sup>-2</sup> s<sup>-1</sup> irradiance from cool white fluorescent lamps.

The growth of *M. aeruginosa* was tested in natural seawater (control), in phosphorus limitation condition (DIN : DIP = 100), in nutrient balanced (DIN : DIP = 16), as well as in nitrogen limitation (DIN : DIP = 3) conditions. All experimental treatments were incubated in three replicates. To test the effect of dissolved organic matter (DOM) a parallel set of treatments was established with DOM addition, also in three replicates. For the experimental phosphorus limitation, balanced N and P and nitrogen limitation treatments, pre-autoclaved sterile nutrient solutions of NaNO<sub>3</sub> and/or KH<sub>2</sub>PO<sub>4</sub> and 0.22 µm filtered DOM were aseptically added (Table 1). Cultures were grown in filter-sterilized (0.22 µm) seawater (4 L) in 5-L polycarbonate pre-autoclaved bottles. Enrichments of Fe-EDTA, metals and vitamins were added. Nutrient additions were made gradually from Day 0 to Day 6. Pure seawater and seawater with DOM addition was used as a control in triplicate.

*M. aeruginosa* growth was monitored daily by *in vivo* fluorescence with a 10 AU Turner Fluorometer (Brand et al. 1981). Estimation of cell number and nutrient concentration was performed on Day 7 (one day after the end of the nutrient addition), Day 9 (end of exponential growth phase), Day 12 and Day 15 (stationary phase). Cell counting was performed until Day 24.

*M. aeruginosa* cell samples (5 mL; Neubaeur cell) were fixed with acid Lugol's solution and counted under a microscope. Chlorophyll *a* was determined spectrophotometrically after filtration through glass microfibre filters (Whatman GF/F), and extraction with pure methanol, at 4 °C, for 1 h in the dark (Jespersen, Christofersen 1987).

Inorganic nutrient concentrations were measured with a Skalar autoanalyzer, following standard procedures for seawater (Valderrama 1995).

Samples for particulate organic carbon (POC) and particulate organic nitrogen (PON) analysis were collected on precombusted glass microfibre filters Whatman GF/F, then decarbonated (HCl fumes, overnight), and analyzed using a CHN analyzer, model 1500 from Carlo Erba. Samples for particulate organic phosphorus (POP) were collected as for

**Table 1.** Concentration of dissolved organic matter (DOM), dissolved inorganic nitrogen (DIN) and dissolved inorganic phosphorus (DIP) added to control (Control + DOM), nutrient balanced (DIN : DIP = 16; (DIN : DIP = 16) + DOM)), phosphorus limited (DIN : DIP = 100; (DIN : DIP = 100) + DOM) and nitrogen limited treatments (DIN : DIP = 3; (DIN : DIP = 3) + DOM), dissolved organic nitrogen – DON; dissolved organic phosphorus – DOP; nitrate –  $NO_3$ ; phosphate –  $PO_4$ .

Treatment	Nutrient	Concentration (µM)
Control	-	-
+ DOM	DON	23.1
	DOP	0.18
DIN: DIP = 100	NO <sub>3</sub>	66.4
	$PO_4$	0.66
(DIN:DIP = 100) + DOM	NO <sub>3</sub>	66.4
	$PO_4$	0.66
	DON	23.1
	DOP	0.18
DIN:DIP=16	NO <sub>3</sub>	66.4
	$PO_4$	4.16
(DIN:DIP = 16) + DOM	NO <sub>3</sub>	66.4
	$PO_4$	4.16
	DON	23.1
	DOP	0.18
DIN: DIP = 3	NO <sub>3</sub>	12.5
	$PO_4$	4.16
(DIN:DIP=3) + DOM	NO <sub>3</sub>	12.8
	$PO_4$	4.16
	DON	23.1
	DOP	0.18

POC and PON. POP concentration was obtained as the  $PO_4$  concentration after persulfate digestion at 120 °C (Pujo-Pay, Raimbault 1994).

DOM extract was obtained in May, June and July 1999 from the River Pärnu water, which flows into the Gulf of Riga. Extraction was made using a tangential flow ultra filter (Benner et al. 1997). Firstly, river water was filtered through 1.2  $\mu$ m and 0.2  $\mu$ m Opticap filter units (Millipore); then, DOM was concentrated with a tangential device, the Prep/scaleTM TFF 6 ft2 cartridge (Millipore). The fraction of U = 10<sup>3</sup> - 10<sup>6</sup> Daltons [Da] [abridged >1000] was obtained and used. Total dissolved nitrogen and phosphorus concentrations were determined after persulfate digestion at 120 °C (Pujo-Pay, Raimbault 1994); DON was calculated as the difference between total nitrogen concentration and DIN (NO<sub>3</sub> + NO<sub>2</sub> + NH<sub>4</sub>), and DOP concentrations was obtained as difference between total phosphorus and PO<sub>4</sub>. The obtained concentrations of DON and DOP in concentrates of DOM were 639  $\mu$ M and 5.14  $\mu$ M, respectively (DON : DOP = 124). The DOM extract contained 49  $\mu$ M NO<sub>3</sub> and 3.37  $\mu$ M PO<sub>4</sub>.

The microcystin content was estimated by protein phosphatase 1A inhibition assay (Ward et al. 1997).

Statistical analysis of variance (ANOVA) was used to test the significance of differences in *M. aeruginosa* growth, biomass increase and nutrient dynamics between experimental conditions

#### Results

The growth of *M. aeruginosa* started without any lag phase in the control and the + DOM conditions. The cell concentration increased until Day 6 in the control, and until Day 14 in the + DOM condition (Fig. 1). With inorganic and organic nutrient additions, cyanobacteria growth occurred according to the typical pattern for a batch cultures: a 1 to 2 day lag phase followed by an exponential-growth phase lasting from Day 3 until Day 10 or Day 11. During the last five days of the experiment the cell concentration decreased. Differences in cell concentration due to the addition of DOM were visible mostly during the stationary phase.

Chlorophyll *a* concentrations started to increase exponentially from Day 1. Maximal chlorophyll *a* concentrations in the control and the + DOM condition were observed already on Day 5, in other conditions with inorganic and organic nutrient additions at Days 6 - 8 (Fig. 2).

As nutrient additions ended on Day 6, the cultures evolved as batch cultures after that time. Nutrient supplies decreased continuously until the end of culturing. In some cases nutrients were taken up rapidly. This occurred for nitrates (NO<sub>3</sub>) in the DIN : DIP-balanced conditions (p < 0.05), where NO<sub>3</sub> concentration already reached <1  $\mu$ M on Day 9 and



**Fig. 1.** Cell number of *Microcystis aeruginosa* versus time (days) in the treatment without inorganic nutrient enrichments (A), treatment with nutrient enrichments in balanced ratio (B), N – limited (C) and P – limited treatment (dissolved inorganic nitrogen – DIN; dissolved inorganic phosphorus – DIP; dissolved organic matter – DOM), vertical bars indicates standard deviation between triplicate.



**Fig. 2.** Chlorophyll *a* concentration of *Microcystis aeruginosa* versus time (days) in the treatment without inorganic nutrient enrichments (A), treatment with nutrient enrichments in balanced ratio (B), N – limited (C) and P – limited treatment (dissolved inorganic nitrogen – DIN; dissolved inorganic phosphorus – DIP; dissolved organic matter – DOM), vertical bars indicates standard deviation between triplicate.

decreased near to detection limit at the end of sampling on Day 15. In contrast, phosphate  $(PO_4)$  concentration decreased slightly in the same condition and large amounts of  $PO_4$  were left unused in the cultures on Day 15. In all treatments, NO<sub>3</sub> and PO<sub>4</sub> concentrations were lowest on Day 15.

In all nutrient-enriched cultures cell number together with particulate organic carbon (POC), particulate organic nitrogen (PON) and particulate organic phosphorus (POP) concentrations continued to increase after the last nutrient addition on Day 6, indicating uptake of N and P and biomass production. For all replicates particulate organic carbon, nitrogen and phosphorus concentration on Days 12 and 15 reached the maximum values. Thus, values recorded on Day 15 were used to determine yields and to estimate the relative cellular contents of carbon, nitrogen and phosphorus.

Nitrogen : phosphorus balanced treatment (DIN : DIP = 16 and DIN : DIP = 16) + DOM)

The highest *M. aeruginosa* numbers and chlorophyll *a* concentrations were observed in the DIN : DIP balanced treatment. Nitrate and phosphate addition in balanced ratio yielded the highest POC, PON and POP increment, while addition of dissolved organic matter had little effect. The biomass increases sustained by DOM were only 7 % for POC, 8 % for PON and 9 % for POP (Table 2).

#### Phosphorus limited treatment (DIN : DIP = 100 and (DIN : DIP = 100) + DOM)

The second highest cyanobacteria cell and chlorophyll *a* concentrations were obtained under P-limitation. Biomass increases obtained by nitrate and phosphate additions

were lower as in the balanced treatment. *M. aeruginosa* growth in phosphorus-limiting conditions did not benefit so much from the addition of DOM; the corresponding value for POC was 14 %; for PON 13 % and 10 % for POP.

### Nitrogen-limiting treatment (DIN : DIP = 3 and (DIN : DIP = 3) + DOM)

Cell number, chlorophyll *a* concentration, POC and PON concentrations were significantly lower (p < 0.05) under N limitation than in nitrogen: phosphorus balanced and phosphorus limited conditions (Fig. 1, 2, Table 2). Cells that were nitrogen-limited seemed to benefit from the addition of DOM; POP concentration in the (DIN : DIP = 3) + DOM treatment represented an increase of 20 % compared to the DIN : DIP = 3 treatment (p < 0.05).

# Unenriched seawater with and without DOM (Control and + DOM)

The *M. aeruginosa* counts in the control and + DOM treatment were lower than in all inorganic nutrient addition treatments (p < 0.05, Fig. 1, 2, Table 2). However the addition of DOM to the unenriched seawater (control) significantly (p < 0.01) increased cyanobacteria *M. aeruginosa* cell number (Fig. 1) and chlorophyll *a* concentration (increase coefficient:  $1.8 \pm 0.3$ , n = 24) in comparison to the control. The increase of cell number was followed by a significant increase of PON and POP concentrations in the +DOM treatment compared to the unenriched control (Table 2). The highest increase by 53 % (p < 0.05) occurred for POP concentration, compared to 31 % (p < 0.05) for PON concentration. We infer that *M. aeruginosa* acquired nitrogen and phosphorus from DOM in conditions where inorganic nutrients were not added, and that DOP of land origin is to some extent a substitute for PO<sub>4</sub> in P-depleted condition.

**Table 2.** Mean particulate organic carbon (POC) concentrations, particulate organic nitrogen (PON), particulate organic phosphorus (POP) and PON : POP ratios of Microcystis aeruginosa recorded on Day 15 in nutrient balanced and nutrient limited experimental treatments (dissolved inorganic nitrogen – DIN; dissolved inorganic phosphorus – DIP; dissolved organic matter – DOM). Number after ( $\pm$ ) indicates standard deviation between three replicates. Numbers within parentheses show relative increase of POC, PON and POP concentrations sustained by the addition of dissolved organic matter, in percents.

Treatment	POC (µM)	PON (µM)	POP (µM)	<b>PON : POP</b>
Control	$250 \pm 23$	$12.1 \pm 2.5$	$0.53\pm0.23$	$22.8\pm8.3$
+ DOM	$340 \pm 37$	$17.5 \pm 1.4$	$1.12\pm0.09$	$15.6 \pm 1.7$
	(26 %)	(31 %)	(53 %)	
DIN : DIP = 100	$700 \pm 32$	$54.9 \pm 1.5$	$1.28\pm0.02$	$42.8\pm1.0$
(DIN : DIP = 100) + DOM	$820 \pm 21$	$62.8 \pm 1.5$	$1.43\pm0.03$	$43.9 \pm 1.3$
	(14 %)	(13 %)	(10 %)	
DIN: DIP = 16	$790 \pm 39$	$61.2 \pm 0.2$	$3.04\pm0.11$	$20.1\pm0.9$
(DIN:DIP = 16) + DOM	$850\pm38$	$66.9 \pm 1.5$	$3.33\pm0.17$	$20.2\pm1.4$
	(7%)	(8 %)	(9 %)	
DIN: DIP = 3	$450 \pm 19$	$23.7 \pm 2.6$	$3.02\pm0.09$	$7.8 \pm 0.8$
(DIN:DIP=3) + DOM	$610 \pm 12$	$26.8\pm0.9$	$3.79\pm0.25$	$7.1 \pm 0.7$
	(26 %)	(12 %)	(20 %)	

### Discussion

M. aeruginosa reached the highest cell and chlorophyll a concentrations in the nitrogen and phosphorus balanced treatment (DIN : DIP = 16 and (DIN : DIP = 16) + DOM). Since the time when Redfield (1934) observed that NO<sub>2</sub>-N and PO<sub>4</sub>-P are taken up by phytoplankton at a constant atomic ratio of 16:1 and Fleming (1940) pointed out that this value is also that of the elemental composition of phytoplankton, it has been commonly accepted that values >16 would reflect P-limited and values < 16 would reflect N-limited cell contents. PON : POP values recorded in our two DIN : DIP balanced conditions were somewhat higher (approximately 20) than the expected Redfield's value - 16. It has been reported that chemical contents of cells vary between species. For example Sakshaug and Holm-Hansen (1977) concluded that the point of change for P-deficiency in cells differs between species, for example, for Skeletonema costatum PON : POP = 23. We assume that cells of *M. aeruginosa* in DIN : DIP = 16 conditions were neither N- nor P-limited. The DIN and DIP supply in balanced ratio (16) is the main determinative factor for development of Microcystis aeruginosa, promoting high cell biomass and chlorophyll a content. The addition of DOM did not affect the cell N : P ratio in the presence of NO<sub>2</sub> and PO<sub>4</sub> additions.

In the phosphorus limited treatments (DIN : DIP = 100 and (DIN : DIP = 100) + DOM) the *M. aeruginosa* count was lower than in the DIN : DIP balanced treatment. As the PON : POP ratios of the cells grown in the phosphorus-limiting treatments were  $42.8 \pm 1.0$  and  $43.9 \pm 1.3$  (Table 2) we infer that the cells in both treatments were clearly phosphorus-limited. The DOM addition raised *M. aeruginosa* biomass and cell POP and PON content, while the PON : POP ratio became more P limited, showing that DOP could not compensate the P-deficiency. Similarly Panosso and Gràneli (2000) showed that DIP-limited *Nodularia spumigena* was unable to use DOM as a source of phosphorus.

*M. aeruginosa* cells grown in the nitrogen-limiting treatment (DIN : DIP = 3 and (DIN : DIP = 3) + DOM) clearly showed nitrogen-limitation, as cell PON : POP was  $7.8 \pm 0.8$  and  $7.1 \pm 0.7$ . Hillebrand and Sommer (1999) considered that a N : P < 13 in microalgae indicates nitrogen limitation. In contrast, Panosso and Gràneli (2000) observed than DON can provide cyanobacteria cells with nitrogen and increase the yield of *Nodularia spumigena* under conditions of N deficiency. These differing results might have originated from different protocols – extremely severe N and P limitation (DIN : DIP 0.02 and 643) and high amount of DOM added, while in our experiment DOM was added at in-situ like concentrations.

The addition of DOM to unenriched seawater promoted a significant decrease of cell PON : POP ratio by 32 %, changing from a P-limited (PON : POP = 22.8) to a N : P-balanced (PON : POP = 15.6) content (Table 2). As in this treatment the DON : DOP ratio in the medium was 9.2, we infer that cells took up proportionally more organic phosphorus than nitrogen. This idea is supported by the literature, as it is known that several cyanobacteria can acquire phosphorus from organic compounds, and *M. aeruginosa* has large capacity for organic phosphorus mineralization Giraudet et al. (1997). Therefore, we conclude that in the treatment where inorganic phosphorus was not added, *M. aeruginosa* mostly took up dissolved organic phosphorus from river DOM.

The utilization of dissolved organic matter (DOM) in aquatic ecosystems was long thought to be the purview of heterotrophic microorganisms, till Paerl (1991) asked a

fundamental question: how can relatively high numbers of actively growing picoplankton survive and periodically flourish in dissolved inorganic nitrogen-depleted waters? He was the first who showed the light-stimulated incorporation of amino acids by cyanobacteria *Synechococcus* species. Other studies have shown that algal biomass and alkaline phosphatase activity are significantly higher in the presence of riverine humic substances than in nitrate-enriched controls (Carlsson et al. 1993). DOM contributes to a biomass increase of cyanobacteria (Ponosso, Granéli 2000) and iron bound in DOM ensures the iron demands (Stolte et al. 2006). According to Maestrini et al. (1999) the DOM from Daugava River promoted *M. aeruginosa* yield to 38 % of that sustained by optimal nitrate concentration.

Our experiments show that the growth of *M. aeruginosa* was promoted by DOM, but since our cyanobacteria cultures were not bacteria-free, we do not infer that *M. aeruginosa* directly obtained all used N and P from DOM. Bacteria associated with *M. aeruginosa* are involved in ammonification and nitrification (Purvina, unpublished data), but still the mechanisms of DOM and particularly DOP uptake are not unequivocal. DOM enhances cyanobacteria biomass production via algal mixotrophy together with bacterial degradation and photochemical modification of marine humic substances. Altogether, the best growth of *M. aeruginosa* isolated from the Gulf of Riga was achieved by nitrate and phosphate additions in the DIN : DIP balanced (16) treatment. The addition of DOM to the control and the different DIN + DIP conditions increased cell number and raised cellular POC, PON and POP concentrations. In inorganic nutrient depleted conditions, *M. aeruginosa* can acquire nitrogen and phosphorus from riverine dissolved organic matter.

#### Acknowledgements

This study was supported by the European Community Project «DOMTOX» (MAS3-CT97-0149) and ESF. Especial thanks is given to M. Pfeifere (LHEI, Dep. of Experimental Hydrobiology) for *M. aeruginosa* culture isolation and to C. Ward (Department of Biological Sciences, University of Dundee) for microcystin analyses.

#### References

- Balode M., Purina I. 1996. Harmful phytoplankton in the Gulf of Riga (the Baltic Sea). In: Yasumoto T., Oshima Y., Fukuyo Y. (eds) *Harmful and Toxic Algal Blooms*. Intergovernmental Oceanographic Commission of UNESCO, pp. 69–72.
- Benner R., Biddanda B., Black B., McCarthy M. 1997. Abundance, size distribution, and stable carbon and nitrogen isotopic compositions of marine organic matter isolated by tangential-flow ultrafiltration. *Mar. Chem.* 57: 243–263.
- Brand L.E., Guillard R.R.L., Murphy L.S. 1981. A method for the rapid and precise determination of acclimated phytoplankton reproduction rates. *J. Plankton. Res.* 3: 193–201.
- Carlsson P., Segatto Z.A., Granéli E. 1993. Nitrogen bound to humic matter of terrestrial origin a nitrogen pool for coastal phytoplankton? *Mar. Ecol. Progr. Ser.* 97: 105–116.
- Dionisio-Pires L.M., Van Donk E. 2002. Comparing grazing by *Dreissena polymorpha* on phytoplankton in the presence of toxic and non-toxic cyanobacteria. *Freshwater Biol.* 47: 1855–1865.
- Dunton P.G., Walsby A.E. 2005. The diameter and critical colapse pressure in gas vesicles of *Microcystis* are correlated with GvpCs of different length. *FEMS Microbiol. Ecol.* 247: 37–43.
- Fleming R.H. 1940. The composition of plankton and units for reporting populations and production. *Proceedings of the 6<sup>th</sup> Pacific Scientific Congress*, California, 1939, pp. 535–540.

- Giraudet H., Berthon J.L., Buisson B. 1997. A comparison of the daily alkaline phosphatase activity of a cyanobacterium (*Microcystis aeruginosa*) and a diatom (*Synedra capitata*). C. R. Acad. Sci. *III*1320: 451–458.
- Granéli E., Johansson N., Panosso, R. 1998. Cellular toxin contents in relation to nutrient conditions for different groups of phycotoxins. In: Reguera B., Blanco J., Fernandez M.L., Wyatt T. (eds) *Harmful Algae*. Xunta de Galicia and Intergovern. Oceanographic Commission of UNESCO, pp. 321–324
- Guillard R.R.L., Ryther J.H. 1962. Studies on marine planktonic diatoms I. *Cyclotella nana* Hustedt and *Detonula confervacea* (Cleve) Gran. *Can. J. Microbiol.* 8: 229–239.
- HELCOM 2007. *Climate Change in the Baltic Sea Area*. HELCOM Thematic Assessment Report on Climate Change in the Baltic Sea Area. Helsinki Commission. 49 p.
- Hillebrand H., Sommer U. 1999. The nutrient stoichiometry of benthic microalgal growth: Redfield proportions are optimal. *Limnol. Oceanogr.* 44: 440–446.
- Humborg C., Conley D.J., Rahm L., Wulff F., Cociasu A., Ittekkot V. 2000. Silicon retention in river basins: Far-reaching effects on biogeochemistry and aquatic food webs in coastal marine environments. *Ambio* 29: 45–50.
- Jespersen A.M., Christofersen K., 1987. Measuremnets of chlorophyll-a from phytoplankton using ethanol as extraction solvent. *Arch. Hydrobiol.* 109: 445–454.
- Juhel G., Davenport J., O'Halloran J., Culloty S., Ramsay R., James K., Furey A., Allis O. 2006. Pseudodiarrhoea in zebra mussels *Dreissena polymorpha* (Pallas) exposed to microcystins. *J. Exp. Biol.* 209: 810–816.
- Kardinaal W.E.A., Janse I., Agterveld M.K., Meima M., Snoek J., Mur L.R., Huisman J., Zwart G., Visser P.M. 2007a. Microcystis genotype succession in relation to microcystin concentrations in freshwater lakles. *Aquat. Microb. Ecol.* 48: 1–12.
- Kardinaal W.E.A., Tonk L., Janse I., Hol S., Slot P., Huisman J., Visser P.M. 2007b. Competition for light between toxic and non-toxic strains of the harmful cyanobacterium *Microcystis. Appl. Environ. Microbiol.* 73: 2939–2946.
- Kļaviņš M., Rodionovs V. 2007. Long-term changes of river discharge regime in Latvia. In: Kļaviņš M. (ed) *Climate Change in Latvia*. University of Latvia, Riga, pp. 21–36.
- Kononen K., Sellner K.G. 1995. Toxic cyanobacteria blooms in marine, estuarine and coastal ecosystems. In: Lassus P., Arzul G., Erard-Le Denn E., Gentien P., Marcaillou-Le Baut C. (eds) *Harmful Marine Algal Blooms*, 6<sup>th</sup> International Conference on Toxic Marine Phytoplankton, Nantes, France. Lavoisier, Paris, pp. 858–860.
- Kurmayer R., Dittmann E., Fastner J., Chorus I. 2002. Diversity of microcystin genes within a population of the toxic cyanobacterium *Microcystis* spp., in Lake Wannsee (Berlin, Germany). *Microbial Ecol.* 43: 107–118.
- Lürling M. 2003. Daphnia growth on microcystin-producing and microcystin-free *Microcystis aeruginosa* in different mixtures with the green alga *Scenedesmus obliguus*. *Limnol. Oceanogr.* 48: 2214–2220.
- Maestrini S.Y, Balode M., Béchemin C., Purina I. 1999. Nitrogenous organic substances as potential nitrogen sources, for summer phytoplankton in the Gulf of Riga, eastern Baltic Sea. *Plankton Biol. Ecol.* 46: 8–17.
- Paerl H.W. 1988. Nuisance phytoplankton blooms in coastal, estuarine, and inland waters. *Limnol. Oceanogr.* 33: 823–847.
- Paerl H.W. 1991. Ecophysiological and trophic implication of light-stimulated amino acid utilization in marine picoplankton. *Appl. Environ. Microbiol.* 57: 473–479.
- Panosso R., Granéli E. 2000. Effects of dissolved organic matter on the growth of *Nodularia spumigena* (Cyanophyceae) cultivated under N or P deficiency. *Mar. Biol.* 136: 331–336.
- Pujo-Pay M., Raimbault P. 1994. Improvement of the wet-oxidation procedure for simultaneous determination of particulate organic nitrogen and phosphorus collected on filters. *Mar. Ecol. Progr. Ser.* 105: 203–207.

- Rahm L., Conley D., Sandén P., Wulff F., Stalnacke P. 1996. Time series analysis of nutrient inputs to the Baltic Sea and changing DSi:DIN ratios. *Mar. Ecol. Progr. Ser.* 130: 221–228.
- Rantajärvi E., Gran V., Hällfors S., Olsonen R. 1997. Effects of environmental factors on the phytoplankton community in the Gulf of Finland unattended high frequency measurements and multivariate analyses. *Hydrobiologia* 363: 1–3.
- Redfield A.C. 1934. On the proportions of organic derivatives in sea water and their relation to the composition of plankton. In: Daniel R.J. (ed) *James Johnstone Memorial Volume*, The University Press, Liverpool, pp. 176-192.
- Sakshaug E., Holm-Hansen O. 1977. Chemical composition of Skeletonema costatum (Grev.) Cleve and Pavlova (Monochrysis) lutheri (Droop) Green as a function of nitrate-, phosphate-, and ironlimited growth. J. Exp. Mar. Biol. Ecol. 29: 1–34.
- Seppälä J., Balode M., 1999 Spatial distribution of phytoplankton in the Gulf of Riga during spring and summer stages. J. Marine Syst. 23: 51–67
- Stolte W., Balode M., Carlsson P., Grzebyk D., Janson S., Lips I., Panosso R., Ward C.J., Granéli E. 2007. Stimulation of nitrogen-fixing cyanobacteria in a Baltic Sea plankton community by landderived organic matter or iron addition. *Mar. Ecol. Progr. Ser.* 327: 71–82.
- Valderrama J.C., 1995. Methods of nutrient analysis. In: Hallegraeff G.M., Anderson D.M., Cembella A.D. (eds) *Manual on Harmful Marine Microalgae*, IOC Manuals and Guides N° 33. UNESCO, Paris, pp. 251–282.
- Ward C.J., Beattie K.A., Lee E.Y.C., Codd G.A. 1997. Colorimetric protein phosphatase inhibition assay of laboratory strains and natural blooms of cyanobacteria: comparison with high-prformance liquid chromatographic analysis for microcystins. *Microbial Ecol.* 153: 465–473.
- Yentsch C.S., Menzel D.W. 1963. A method for the determination of phytoplankton and pheophytin by fluorescence. *Deep-Sea Res.* 10: 221–231.

# Neorganisko barības vielu un izšķīdušās organiskās vielas ietekme uz Rīgas līča cianobaktērijas *Microcyctis aeruginosa* augšanu

# Santa Purviņa<sup>1\*</sup>, Christian Bechemin<sup>2</sup>, Maija Balode<sup>1</sup>, Daniel Grzebyk<sup>2</sup>, Serge Y. Maestrini<sup>2</sup>

<sup>1</sup>Latvijas Hidroekeoloģijas institūts, Daugavgrīvas 8, Rīga LV-1048, Latvija <sup>2</sup>Centre de Recherche en Ecologie Marine et Aquaculture de L'Houmeau (CNRS-IFREMER), B.P. 5, F-17137 L'Houmeau, Francija

\*Korespondējošais autors, E-pasts: santa.purvina@lhei.lv

# Kopsavilkums

Pētīja barības vielu ietekmi N-limitējošos, P-limitējošos un DIN : DIP līdzsvarotos apstākļos, ar un bez sauszemes izcelsmes izšķīdušās organiskās vielas (DOM) klātbūtnes, uz cianobaktērijas *Microcystis aeruginosa*, kas izolēta no Rīgas līča, augšanu un toksīna saturu. Izšķīdušā neorganiskā slāpekļa (DIN) un izšķīdušā neorganiskā fosfora (DIP) pievienošana līdzsvarotā attiecībā DIN : DIP = 16 ir galvenais *M. aeruginosa* attīstību noteicošais faktors, kas veicina šūnu biomasas, hlorofila *a* un daļiņveida organiskās vielas pieaugumu. Vidē, kas bija bagātināta tikai ar DOM, *M. aeruginosa* izmantoja 31 % no pieejamā izšķīdušā organiskā slāpekļa un 53 % no izšķīdušā organiskā fosfora. Eksperimenta apstākļi neveicināja cianobaktērijas toksiskuma palielināšanos. Eksperiments parādīja, ka *M. aeruginosa* attīstību veicina paaugstināta neorganisko barības vielu un izšķīdušās organiskā svielas ieplūde iesāļajās Baltijas jūras piekrastes ekosistēmās.